

# Carcinogenic Polycyclic Aromatic Hydrocarbons in Urban Air Particulate Matter

A. Szabó Nagy, J. Szabó, Zs. Csanádi, J. Erdős

**Abstract**—An assessment of the air quality of Győr (Hungary) was performed by determining the ambient concentrations of PM10-bound carcinogenic polycyclic aromatic hydrocarbons (cPAHs) in different seasons. A high volume sampler was used for the collection of ambient aerosol particles, and the associated cPAH compounds (benzo[a]pyrene (BaP), benzo[a]anthracene, benzofluoranthene isomers, indeno[123-cd]pyrene and dibenzo[ah]anthracene) were analyzed by a gas chromatographic method. Higher mean concentrations of total cPAHs were detected in samples collected in winter (9.62 ng/m<sup>3</sup>) and autumn (2.69 ng/m<sup>3</sup>) compared to spring (1.05 ng/m<sup>3</sup>) and summer (0.21 ng/m<sup>3</sup>). The calculated BaP toxic equivalent concentrations have also reflected that the local population appears to be exposed to significantly higher cancer risk in the heating seasons. Moreover, the concentration levels of cPAHs determined in this study were compared to other Hungarian urban sites.

**Keywords**—Air, carcinogenic, PAH, PM10.

## I. INTRODUCTION

POLYCYCLIC aromatic hydrocarbons (PAHs) are a class of organic compounds that consist of two or more fused aromatic rings without heteroatoms. The United States Environmental Protection Agency (US EPA) has selected 16 of these compounds (priority PAHs) for measurement in various environmental samples [1]-[3]. PAHs are generated by natural and anthropogenic sources [2]-[4]. Natural emissions come from forest fires, biosynthesis of some plants and volcanic eruptions. The incomplete burning of fossil fuels and organic materials, emissions of motor vehicles, power generation and industries are primarily responsible for the anthropogenic emissions of PAHs [2]-[4].

PAHs exist in the atmosphere in both vapor and particulate phase. The particle-associated PAHs are considered to be more harmful to human health because they are inhalable and ingestible and can deposit in the human respiratory system [5], [6]. The analysis of atmospheric particles with an aerodynamic diameter <10 μm (PM10) or less (e.g. PM2.5 and PM1) from urban regions has become important, considering the biological effects and potential health hazards they can impose. Particulate matter comprises several compounds including PAHs and their derivatives which display mutagenic and carcinogenic properties.

A. Szabó Nagy, J. Szabó and Zs. Csanádi are with the Physics and Chemistry Department, Széchenyi István University, Győr, H 9026 Hungary (phone: 36-96-503-168; fax: 36-96-613-558; e-mail: nszaboa@sze.hu; jszabo@sze.hu; csanzs@sze.hu).

J. Erdős is with the Government Office for Győr-Moson-Sopron County, Environmental Protection Laboratory, Győr, H 9028 Hungary (e-mail: erdos.jozsef@gyor.gov.hu).

In this study, the seasonal distributions of several US EPA probable or class B2 human carcinogen PAHs (benzo[a]pyrene (BaP), benzo[a]anthracene (BaA), benzo[b]fluoranthene, benzo[k]fluoranthene, benzo[j]fluoranthene, indeno[123-cd]pyrene (IND) and dibenzo[ah]anthracene (DahA)) associated with PM10 were assessed in Győr, Hungary. The levels of carcinogenic PAHs (cPAHs) determined in this study was compared to other Hungarian urban sites and a carcinogenic risk assessment model was also obtained for the concentration data.

## II. MATERIALS AND METHODS

### A. Sampling

A total of 56 PM10 aerosol samples were collected at the monitoring station of Győr in the year of 2014. Győr is the most important city in northwest Hungary, and one of the seven main regional centers of the country [7], [8]. The location of the city is shown in Fig. 1. The number of 24-hour sampling days was 14 in every season. A Digital High Volume sampler DHA-80 (Digital Elektronik AG, Switzerland) was used for the sampling [9].

### B. Analytical Procedure

The ultrasonic liquid-solid extraction of the filter sample (Advantec QR-100 quartz fibre) and the PAH analysis (BaP, BaA, IND, DahA and the sum of the three benzofluoranthene isomers (BF)) were conducted in accordance with the Hungarian standard method procedure [10]. The method of the analytical procedure and the gas chromatographic system were given in detail in our previous works [7], [8].

### C. BaP Equivalent Concentrations

The carcinogenic risk of PAHs mixture is often expressed by its BaP equivalent concentration (BaPE). The BaPE of atmospheric PAHs was calculated from (1) [7], [8], [11]:

$$BaPE = BaA \times 0.06 + BF \times 0.07 + BaP + DahA \times 0.6 + IND \times 0.08 \quad (1)$$

### D. Cancer Risk Estimates

According to US EPA [12], a human exposure depends on chronic daily intake (CDI) of every single contaminant inhaled by the receptor. CDI for carcinogenic substances is called life averaged daily dose (LADD) [13]. The LADD value (mg/kg-day) can be derived from (2) [12]-[15]:

$$LADD = CA \times IR \times ET \times EF \times ED / BW \times AT \quad (2)$$

where CA is a compound concentration (mg/m<sup>3</sup>), IR

(inhalation rate) is a breathing rate ( $\text{m}^3/\text{h}$ ), EF (exposure frequency) is a number of exposures per year, ED (exposure duration) is a duration of exposure in years, ET (exposure time) is a number of hours per exposure, BW (body weight) is a default weight of the receptor body (kg), and AT (averaging time) is an average exposure extent over a lifetime (25,550 days, equal to 70 years for carcinogens) [13]–[15]. In this study, the CA was used for BaPE concentration.



Fig. 1 The location of Győr and some Hungarian monitoring stations

Appropriate default exposure parameters were obtained by an updated US EPA recommendation for resident adult ( $\text{IR} = 1 \text{ m}^3/\text{h}$ , because inhalation rate is no longer used in (2),  $\text{ET} = 24 \text{ h/day}$ ,  $\text{EF} = 350 \text{ days/year}$ ,  $\text{ED} = 20 \text{ years}$ ,  $\text{BW} = 80 \text{ kg}$ ) [16].

Human health risk related to contaminated air depends on the extent exposure as well as the toxic effects on chemicals. The incremental lifetime cancer risk (ILCR) was calculated from (3) [12], [14], [15]:

$$\text{ILCR} = \text{LADD} \times \text{SF} \quad (3)$$

A slope factor (SF) is used to estimate an upper-bound probability of the individual developing a cancer as result of the lifetime exposure to certain level of potential carcinogen. SF for BaP (7.3 per  $\text{mg}/\text{kg}\cdot\text{day}$ ) was obtained from US EPA [17].

### III. RESULTS AND DISCUSSION

#### A. PM10 Concentration Level

The mean PM10 concentration was  $23.04 \mu\text{g}/\text{m}^3$  at the monitoring station of Győr, which is about two times less than the annual limit value of  $40 \mu\text{g}/\text{m}^3$  (Tables I and II). On average, a good air quality index was identified for PM10. Only 3.13% of the total samples exceeded the daily limit value of  $50 \mu\text{g}/\text{m}^3$ .

#### B. Occurrence and Seasonal Distribution of PAHs

All PAHs examined in this study were identified in the aerosol samples. The mean concentrations were in order of  $\text{BF} > \text{BaP} > \text{IND} > \text{BaA} > \text{DahA}$  (Table II). The seasonal variation of individual PAHs and total cPAHs concentrations are shown in Figs. 2 and 3.

TABLE I  
AIR QUALITY LIMIT OR TARGET VALUES AND QUALITY INDEX USED IN HUNGARY

Parameter	PM10		BaP	
	Daily ( $\mu\text{g}/\text{m}^3$ )	Annual ( $\mu\text{g}/\text{m}^3$ )	Daily ( $\text{ng}/\text{m}^3$ )	Annual ( $\text{ng}/\text{m}^3$ )
EU limit or target value [5]	50 <sup>a</sup>	40	–	1
Hungarian limit value [18]	50 <sup>a</sup>	40	1	0.12
WHO guideline value [6]	50 <sup>b</sup>	20	–	0.12 <sup>c</sup>
Quality index <sup>d</sup>				
1. Excellent	0–20	0–16	0–0.4	0–0.048
2. Good	20–40	16–32	0.4–0.8	0.048–0.096
3. Acceptable	40–50	32–40	0.8–1	0.096–0.12
4. Polluted	50–100	40–80	1–2	0.12–0.24
5. Heavily polluted	100≤	80≤	2≤	0.24≤

<sup>a</sup>Not to be exceeded on more than 35 days/year.

<sup>b</sup>99th percentile (3 days/year).

<sup>c</sup>As the WHO has not set AQG for BaP, the reference level in was estimated assuming WHO unit risk for lung cancer for PAH mixtures, and an acceptable risk of additional lifetime cancer risk of approximately  $1 \cdot 10^{-5}$ .

<sup>d</sup>Based on the Hungarian limit values (excellent: 0–40 %; good: 40–80 %; acceptable: 80–100 %; polluted: 100–200 %; heavily polluted:  $\geq 200$  %).

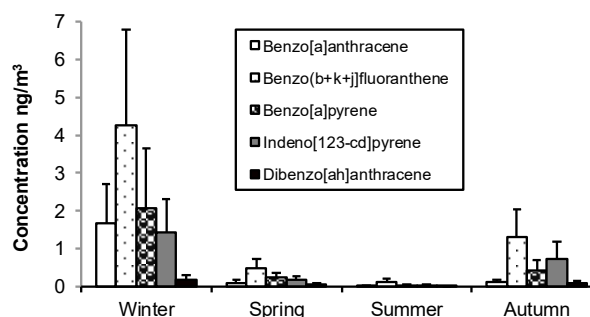


Fig. 2 Seasonal variation of individual PAHs at the monitoring station of Győr

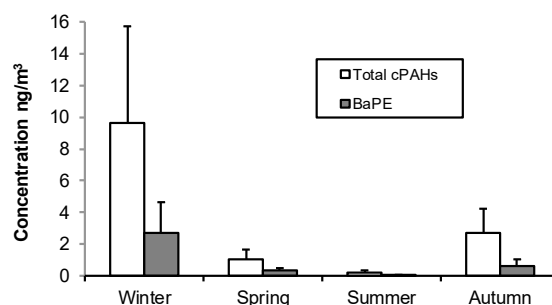


Fig. 3 Seasonal variation of total cPAHs and BaPE at the monitoring station of Győr

The total cPAHs mean concentrations were  $1.05 \text{ ng}/\text{m}^3$  and  $0.21 \text{ ng}/\text{m}^3$  in spring and summer, respectively. The elevated concentrations of particle-associated PAHs and total cPAHs in winter ( $9.62 \text{ ng}/\text{m}^3$ ) and autumn ( $2.69 \text{ ng}/\text{m}^3$ ) can be attributed to the increased emissions from heating sources, the transformation of gas/particle partitioning toward the particulate-phase induced by low temperature, the reduced photochemical degradation of some PAHs by solar radiation, and the lower dispersion due to meteorological conditions prevailing in winter [3], [15].

The seasonal profiles of particle-associated PAHs (percent

contribution of each PAH compound to total cPAHs) are illustrated in Fig. 4. The sum of BF isomers accounted the highest percentage contribution in all seasons (45–61 %). The DahA was insignificant (1–5 %). The contribution of BaP and

BaA to total cPAHs was higher in winter and spring than summer or autumn. However, the highest percentage of IND (27%) was measured in autumn.

TABLE II  
PM10 AND PM10-BOUND PAH CONCENTRATION LEVELS FOR THE HUNGARIAN URBAN SITES IN 2014 AS WELL AS THE BACKGROUND SITE [16]

Sites	Maximum and mean concentrations					Percentage of samples exceeded the limit value (%) <sup>a</sup>		Mean percentage contribution of BaP to PM10 (%)	
	PM10 (µg/m <sup>3</sup> )	PAH compounds (ng/m <sup>3</sup> )					PM10		BaP
Győr	76.60	4.02	9.73	5.87	0.49	3.37	3.13	17.86	2.95·10 <sup>-3</sup>
	23.04	0.48	1.55	0.68	0.08	0.60			
Miskolc	154.26	8.20	12.43	5.76	1.73	4.60	42.53	50.0	3.42·10 <sup>-3</sup>
	51.98	2.07	4.60	1.78	0.46	1.59			
Nyíregyháza	99.70	10.76	28.96	13.56	2.23	15.96	23.21	48.21	5.31·10 <sup>-3</sup>
	37.46	1.45	4.94	1.99	0.32	2.46			
Debrecen	74.0	17.63	31.22	20.01	3.57	21.53	12.73	49.09	1.15·10 <sup>-2</sup>
	27.81	2.66	6.41	3.19	0.46	3.18			
Békéscsaba	76.50	8.97	24.04	14.58	1.40	10.29	8.93	44.64	6.64·10 <sup>-3</sup>
	26.18	1.24	4.14	1.74	0.26	1.44			
Tapolca	64.5	1.0	3.90	1.60	0.30	1.20	3.57	8.93	1.57·10 <sup>-3</sup>
	22.25	0.18	0.92	0.35	0.07	0.35			
Dunaujváros	125.6	11.50	42.40	14.20	4.70	1.20	14.29	27.78	3.98·10 <sup>-3</sup>
	30.62	0.88	3.50	1.22	0.34	0.35			
Veszprém	82.80	3.70	10.40	4.40	0.80	13.10	16.07	42.86	4.06·10 <sup>-3</sup>
	28.07	0.93	2.66	1.14	0.19	1.13			
Kecskemét	64.30	6.36	12.64	6.59	ND	7.7	12.50	33.93	3.46·10 <sup>-3</sup>
	27.14	0.54	2.35	0.94		0.94			
Szombathely <sup>b</sup>	62.10	4.85	14.64	11.41	4.85	6.13	5.10	38.78	5.64·10 <sup>-3</sup>
	22.51	0.74	2.56	1.27	0.31	1.67			
Mosonmagyaróvár	71.20	4.61	9.95	7.12	0.65	4.65	1.79	32.14	4.11·10 <sup>-3</sup>
	22.16	0.59	1.82	0.91	0.10	0.73			
Szentgotthárd	64.06	6.66	13.60	8.03	0.97	6.69	3.57	33.93	7.42·10 <sup>-3</sup>
	18.34	0.99	2.76	1.36	0.18	1.39			
Esztergom	64.80	5.76	13.53	7.72	0.85	5.76	5.36	23.21	3.85·10 <sup>-3</sup>
	22.31	0.54	1.83	0.86	0.11	0.77			
Budapest <sup>b</sup>	99.10	–	–	9.44	–	–	16.22	44.14	3.60·10 <sup>-3</sup>
	33.91			1.22					
Pécs	95.06	–	–	5.62	–	–	9.52	15.0	2.58·10 <sup>-3</sup>
	31.01			0.80					
Százhalombatta	74.21	–	–	6.56	–	–	16.07	41.07	3.70·10 <sup>-3</sup>
	31.08			1.15					
K-pusztá (background)	54.0	2.45	7.10	3.30	0.64	4.22	3.64	36.36	4.24·10 <sup>-3</sup>
	18.89	0.48	1.97	0.80	0.12	1.02			

<sup>a</sup>Hungarian daily limit values (see Table II).

<sup>b</sup>Based on data of 2 monitoring sites.

– = No data

### C. Carcinogenic Risk Assessment

BaPE concentrations calculated based on the individual PAH concentrations were frequently utilized to assess the potential health risk associated with exposure to PAHs. In this study, the mean value of BaPE for Győr was 0.91 ng/m<sup>3</sup>. However, it had shown similar patterns of seasonal distributions with total PAHs (Fig. 3). The BaPE in winter (2.69 ng/m<sup>3</sup>) was nearly 4 and 8 times higher than that in autumn (0.64 ng/m<sup>3</sup>) and spring (0.32 ng/m<sup>3</sup>), indicating that the heating increased the particle-associated PAHs exposure risk substantially. The BaPE concentration in summer was insignificant (0.04 ng/m<sup>3</sup>) due to the intensive sunlight for degradation [19].

The comparison of the BaPE concentrations observed for Győr with other cities is shown in Fig. 5. It was found that the highest BaPE levels were identified mainly in the eastern part of Hungary. Among the individual PAH compounds, it is evident that the concentration of BaP at the different sites

presents certain risk. The mean and maximum BaP concentrations are summarized in Table II. The concentrations of BaP were often higher than the Hungarian daily limit value of 1 ng/m<sup>3</sup> (Table I). Also, the annual mean BaP concentration levels were above the equal EU annual target value in most Hungarian cities. Heavily polluted annual air quality for BaP (> 2 ng/m<sup>3</sup>) was obtained for only one urban site located in the northern region of the country. For Győr, a good air quality of BaP was indicated based on the EU mean target value of 1 ng/m<sup>3</sup>. The mean percentage contributions of BaP to PM10 (%) are also shown in Table II. The range of BaP contributions of PM10 concentrations was about 0.006 and 0.01% at the different Hungarian sites.

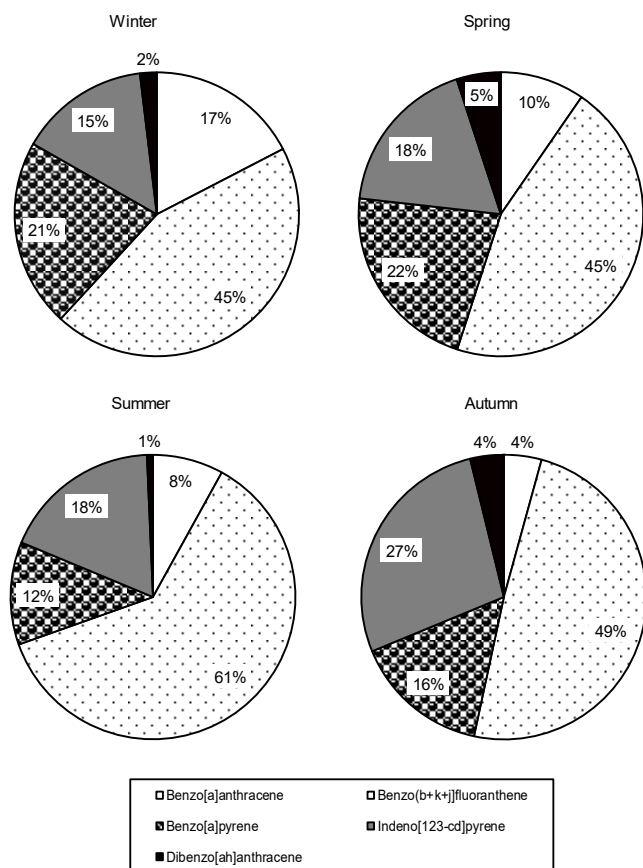


Fig. 4 Composition pattern (%) of PAHs at the monitoring station of Győr

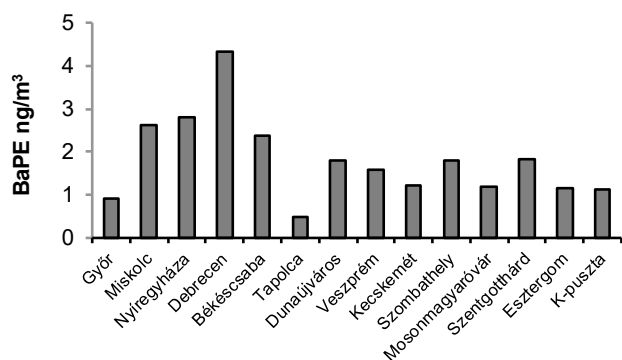


Fig. 5 Mean BaPE concentration levels for the Hungarian urban sites and the background site of K-pusztá

The calculated values of LADD of the examined Hungarian sites, and the associated ILCR values are presented in Table III. LADD values of carcinogenic PAHs by way of inhalation of PM10 were between  $4.07 \times 10^{-8}$  mg/kg-day and  $3.56 \times 10^{-7}$  mg/kg-day. On the basis of LADD values, the total ILCR was calculated to be between  $2.97 \times 10^{-7}$  and  $2.60 \times 10^{-6}$ . These values are close to the acceptable limit of  $10^{-6}$ – $10^{-4}$  as acknowledged by regulatory agencies or not exceeded the value of  $10^{-6}$  [12]. The calculated ILCR value for Győr was  $5.52 \times 10^{-7}$ . Among the Hungarian sites, Győr was on the second tier in terms of ILCR values for PAHs.

TABLE III  
LADD AND ILCR VALUES FOR PAHs OF THE HUNGARIAN URBAN SITES AND THE BACKGROUND SITE OF K-PUSZTA

Sites	LADD (mg/kg·day)	ILCR
Győr	$7.56 \times 10^{-8}$	$5.52 \times 10^{-7}$
Miskolc	$2.16 \times 10^{-7}$	$1.58 \times 10^{-6}$
Nyíregyháza	$2.31 \times 10^{-7}$	$1.69 \times 10^{-6}$
Debrecen	$3.56 \times 10^{-7}$	$2.60 \times 10^{-6}$
Békéscsaba	$1.95 \times 10^{-7}$	$1.43 \times 10^{-6}$
Tapolca	$4.07 \times 10^{-8}$	$2.97 \times 10^{-7}$
Dunaújváros	$1.49 \times 10^{-7}$	$1.09 \times 10^{-6}$
Veszprém	$1.29 \times 10^{-7}$	$9.42 \times 10^{-7}$
Kecskemét	$9.96 \times 10^{-8}$	$7.27 \times 10^{-7}$
Szombathely	$1.49 \times 10^{-7}$	$1.09 \times 10^{-6}$
Mosonmagyaróvár	$9.79 \times 10^{-8}$	$7.15 \times 10^{-7}$
Szentgotthárd	$1.51 \times 10^{-7}$	$1.10 \times 10^{-6}$
Esztergom	$9.44 \times 10^{-8}$	$6.89 \times 10^{-7}$
K-pusztá	$9.21 \times 10^{-8}$	$6.72 \times 10^{-7}$

#### IV. CONCLUSION

A study on atmospheric concentration of cPAHs bound to PM10 was carried out for Győr during the year 2014. A significant seasonally distribution of individual PAHs and total cPAHs was identified. On the basis of the mean ILCR values and the comparison of the data with other Hungarian urban sites or the background site, it is evident that the risk resulting from the exposure to air contaminated by PAHs is acceptable for Győr. However, the seasonally distributions of PAHs and the calculated BaPE concentrations have indicated that the local population appears to be exposed to significantly higher cancer risk in the heating seasons.

#### ACKNOWLEDGMENT

Authors are indebted to István Vass, Bálint Kauker, Zsuzsanna Károly Némethné, Tünde Takács Kovácsné, Lajosné Bakódy and Péter Lautner (Government Office for Győr-Moson-Sopron County, Environmental Protection Laboratory, Hungary) for chemical analyzes, data and site information.

#### REFERENCES

- [1] ATSDR, *Toxicological profile for polycyclic aromatic hydrocarbons*. Agency for Toxic Substances and Disease Registry (ATSDR), Atlanta, U.S. Department of Health and Human Services, 1995.
- [2] K. Srogi, "Monitoring of environmental exposure to polycyclic aromatic hydrocarbons: a review," *Environ. Chem. Lett.*, vol. 5, pp. 169–195, 2007.
- [3] K. Ravindra, R. Sokhi, R. V. Grieken, "Atmospheric polycyclic aromatic hydrocarbons: Source attribution, emission factors and regulation," *Atmos. Environ.*, vol. 42, no. 13, pp. 2895–2921, 2008.
- [4] L. R. Nino, R. J. Torres, A. A. Mozeto, P. S. Fadini, "Using urban streams as drinking water: The potential risk in respect to polycyclic aromatic hydrocarbons (PAHs) content in sediments," *Polycycl. Aromat. Compd.*, vol. 34, no. 5, pp. 518–531, 2014.
- [5] EEA, *Air quality in Europe – 2013 report*. European Environment Agency, Luxembourg, 2013.
- [6] WHO, *Air Quality Guidelines for Europe, global update 2005*. World Health Organization, Regional Office for Europe, Copenhagen, 2005.
- [7] J. Szabó, A. Szabó Nagy, J. Erdős, "Ambient concentrations of PM10, PM10-bound polycyclic aromatic hydrocarbons and heavy metals in an urban site of Győr, Hungary," *Air Qual. Atm. Hlth.*, vol. 8, pp. 229–241, 2015.

- [8] A. Szabó Nagy, J. Szabó, Zs. Csanádi, J. Erdős, "Seasonal variation of polycyclic aromatic hydrocarbons associated with PM10 in Győr, Hungary," *Int. J. Environ. Chem. Ecol. Geol. Geophys. Eng.*, vol. 9, no. 8, pp. 872–876, 2015.
- [9] MSZ EN 12341:2000, *Air quality. Determination of the PM10 fraction of suspended particulate matter*. Reference method and field test procedure to demonstrate reference equivalence of measurement methods, Hungarian Standard Association, Budapest, 2000.
- [10] MSZ EN 15549:2008, *Air quality. Standard method for measurement of the concentration of benzo[a]pyrene in ambient air*. Hungarian Standard Association, Budapest, 2008.
- [11] Y. L. M. A. B. D Limu, D. L. N. T. Lifu, A. B. L. Y. Miti, X. Wang, X. Ding, "Autumn and wintertime polycyclic aromatic hydrocarbons in PM2.5 and PM2.5-10 from Urumqi, China," *Aerosol Air Qual. Res.*, vol. 13, pp. 407–414, 2013.
- [12] US EPA, *Risk assessment guidance for superfund volume I, human health evaluation manual (Part A)*. Office of Emergency and Remedial Response, U.S. Environmental Protection Agency, Washington, 1989.
- [13] P. Bohlin, O. Audy, L. Skrdliková, P. Kukucka, S. Vojta, P. Pribylová, R. Prokes, P. Cupr, J. Klánová, "Evaluation and guidelines for using polyurethane foam (PUF) passive air 6 samplers to assess semi volatile organic compounds (SVOCs) in non-industrial 7 indoor environments," *Environ. Sci. Process Impacts*, vol. 16, no.11, pp. 2617–2626, 2014.
- [14] V. Singla, T. Pachauri, A. Satsangi, K. M. Kumari, A. Lakhani, "Characterization and mutagenicity assessment of PM2.5 and PM10 PAH at Agra, India," *Polycycl. Aromat. Compd.*, vol. 32, pp. 199–220, 2012.
- [15] J. Han, N. Zhang, C. Niu, B. Han, Z. Bai, "Personal exposure of children to particle-associated polycyclic aromatic hydrocarbons in Tianjin, China," *Polycycl. Aromat. Compd.*, vol. 34, pp. 320–342, 2014.
- [16] MRBCA, *Exposure factors*. Missouri risk-based corrective action technical guidance, appendix E, <https://dnr.mo.gov/env/hwp/mrbca/docs/MRBCAExposureFactorsComparison.pdf>.
- [17] US EPA, *Toxicological review of benzo[a]pyrene*. National Center for Environmental Assessment, Office of Research and Development, U.S. Environmental Protection Agency, Washington, 2012.
- [18] OMSZ ÉLFO, *Summary of the OLM PM10 sampling program in 2014*. Reference Centre for Air Quality Protection, Budapest, 2014 (in Hungarian).
- [19] F. Valerio, A. Lazzarotto. "Photochemical degradation of polycyclic aromatic hydrocarbons (PAH) in real and laboratory conditions," *Int. J. Environ. Anal. Chem.*, vol. 23., no. 1–2, pp. 135–151, 1985.